

Concordant Genetic Estimators of Migration Reveal Anthropogenically Enhanced Source-Sink Population Structure in the River Sculpin, *Cottus gobio*

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ABSTRACT

River systems are vulnerable to natural and anthropogenic habitat fragmentation and will often harbor populations deviating markedly from simplified theoretical models. We investigated fine-scale population structure in the sedentary river fish *Cottus gobio* using microsatellites and compared migration estimates from three F_{ST} estimators, a coalescent maximum-likelihood method and Bayesian recent migration analyses. Source-sink structure was evident via asymmetry in migration and genetic diversity with smaller upstream locations emigration biased and larger downstream subpopulations immigration biased. Patterns of isolation by distance suggested that the system was largely, but not entirely, in migration-drift equilibrium, with headwater populations harboring a signal of past colonizations and in some cases also recent population bottlenecks. Up- vs. downstream asymmetry in population structure was partly attributable to the effects of flow direction, but was enhanced by weirs prohibiting compensatory upstream migration. Estimators of migration showed strong correspondence, at least in relative terms, especially if pairwise F_{ST} was used as an indirect index of relative gene flow rather than being translated to Nm . Since true parameter values are unknown in natural systems, comparisons among estimators are important, both to determine confidence in estimates of migration and to validate the performance of different methods.

ESTIMATION of gene flow is a key to understanding and predicting human impacts on natural populations. Traditionally, the principal indirect estimator of gene flow has been F_{ST} , which in WRIGHT's (1951) infinite island model equates explicitly with the effective number of migrants (Nm) via $F_{ST} \approx 1/(4Nm + 1)$. The island model makes many simplifying assumptions, and inference of gene flow from F_{ST} has attracted severe criticism (BOSSART and PROWELL 1998; WHITLOCK and MCCAULEY 1999). Moreover, the impacts of deviations from the island model have primarily considered F_{ST} as an inbreeding (F) coefficient, rather than more widely used estimators of F_{ST} that are based on pairwise allele-frequency variation (NEIGEL 2002). Modern coalescent-based maximum-likelihood estimators of gene flow and Bayesian estimators of recent migration are potentially more accurate because they make fewer assumptions and estimate all parameters simultaneously. While contemporary studies seldom attempt to assess gene flow solely via F_{ST} -based analysis, the accuracy of modern estimators of migration in natural systems exhibiting complex population structures is poorly understood.

Therefore, the application of migration estimators that rely on different assumptions in empirical studies of natural populations is important to give insights into their relative performance and to assess how confidently parameter estimates can be interpreted.

Rivers appear to be particularly problematical systems for the implementation of traditional population genetic analyses. Migration will typically take place only along a dendritic habitat network that, with simple theoretical models, might be best approximated by the linear stepping-stone model, in which migrants are exchanged only between immediate neighbors in a one-dimensional chain of subpopulations (KIMURA and WEISS 1964). Although the relationship between F_{ST} and Nm appears to be quite robust to distance-limited dispersal (SLATKIN and BARTON 1989), near unidirectional flow of water creates a force that promotes movement of organisms along corridors downstream, setting conditions for asymmetric migration and resultant isolation of headwater habitats (FRASER *et al.* 2004). Unequal migration constitutes a major deviation from the symmetry of both the island and stepping-stone models, with both theory and simulations predicting that bias in F_{ST} can ensue (WHITLOCK and MCCAULEY 1999; WAKELEY 2001; WILKINSON-HERBOTS and ETTRIDGE 2004). Isolated populations are expected to be more vulnerable to extinction due to inbreeding or demographic

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effects (SACCHERI *et al.* 1998; MORITA and YAMAMOTO 2002; FAGAN *et al.* 2005). Thus, headwater populations might be prone to show cycles of decline and recovery and might never attain migration-drift equilibrium—a critical assumption for most analytical methods if gene flow estimates are to reflect ongoing, rather than historical, processes. Asymmetry in population size, migration, and the potential for local extinction may be enhanced by human activities that fragment rivers, such as the creation of barriers (*e.g.*, by construction of weirs and culverts) or pollution if these primarily impact active upstream movement. Most previous studies of the effects of human-made barriers have focused on migratory salmonids (MELDGAARD *et al.* 2003; YAMAMOTO *et al.* 2004; VAN HOUTD *et al.* 2005; WOFFORD *et al.* 2005) and little is known about impacts on potentially more vulnerable species with limited dispersal and fecundity. Riverine fishes, particularly sedentary species exhibiting habitat specialization, are thus useful models to examine the following applied and fundamental questions: (1) How are natural populations impacted by human activities? and (2) How congruent are results from different estimators of migration in natural population structures deviating markedly from theoretical models?

The bullhead or river sculpin, *Cottus gobio*, is a widely distributed, but locally endangered, freshwater fish. Bullheads require clean, cool, well-oxygenated waters, with moderate flow rates and at least partial coverage of river beds with reasonably large stones to serve as breeding sites (KNAEPKENS *et al.* 2002a). Such ideal habitat is often patchily distributed along river courses with associated disjunctive distribution of bullheads. Adult movement of *C. gobio* is very limited. Tagging experiments have shown that individuals typically move <10 m during the course of 1 year with only occasional movements of up to 100 m (DOWNHOWER *et al.* 1990; KNAEPKENS *et al.* 2004). Fecundity is moderate, with a maximum of a few hundred eggs deposited at a male's nest site by one to six females (our personal observations). These specialized habitat requirements, coupled with limited dispersal and reproductive potential, suggest that bullhead populations will be sensitive to environmental changes, such as intentional or accidental human activities that fragment river systems. In light of such potential vulnerability, knowledge of the spatial structure and connectivity of subpopulations, and of the resulting consequences for population stability, is important for river management.

In this article we compare several estimators of genetic diversity and migration in a study of sources of habitat fragmentation in *C. gobio*. In spite of multiple deviations from simple theoretical population models, including marked asymmetries in migration and subpopulation sizes, contrasting analytical methods demonstrate an adverse impact of anthropogenic barriers within the river system.

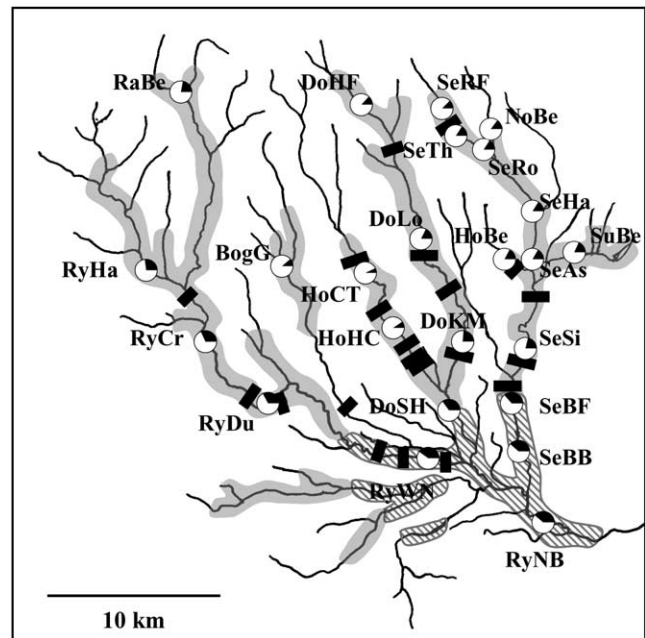


FIGURE 1.—Distribution of sample sites in the River Rye catchment. Shaded areas represent a near-continuous distribution of suitable habitat; hatched areas indicate a more patchy habitat distribution. Latitudes and longitudes: top (of box): 54°24' N; bottom: 54°10' N; left: 01°13' W; right: 00°49' W. Pie charts represent allelic richness at each sample site as a proportion of overall number of alleles.

MATERIALS AND METHODS

Sample sites and geographic information: Our study focused on the River Rye and its tributaries, a river system located in Northeast England. Analysis of long-term monitoring data by the United Kingdom Environment Agency revealed that *C. gobio* is widely distributed across most of the Rye catchment (Figure 1). In the upper parts of the Rye and its tributaries, bullhead habitat is semicontinuous, comprising lengthy stretches interspersed with relatively small sections of unsuitable habitat (≤ 50 m) and *C. gobio* was found consistently and at high density in all potential sample sites. Farther downstream, distributions became patchier and high-density sites were interspersed with unsuitable habitat covering several hundred meters or more. In only one case, however, were bullheads absent for a distance >10 km. Samples of 18–40 bullheads, mostly adults, were taken from 24 sites across the River Rye catchment area by electro-fishing from a river stretch of 50–100 m/site. Sample sites were chosen to cover all main tributaries that contained bullheads and sites of high bullhead densities within tributaries. (Figure 1). A caudal fin clip from each individual was stored in 98% ethanol for genetic analysis. Geographic distances between sample sites were estimated as river distances using Environment Agency databases. As a measure of the relative position of each sample site within the river system, we used “upstream size” (HÄNFLING and BRANDL 1998b), defined as the total length of all suitable river habitats above each site (*i.e.*, catchment length); thus, upstream populations have a lower upstream size than their downstream counterparts. The location of weirs—potential anthropogenic obstructions to bullhead movement—were obtained from Environment Agency databases and maps (D. HOPKINS, personal communication) and confirmed by

personal observations and landowner surveys. A total of 20 weirs were identified (Figure 1) and, although data on their ages were not available, landowners and Environment Agency records confirmed that weirs have been present in the River Rye catchment for at least 200 years. Since the generation time of bullheads is ~1–2 years (NOLTE *et al.* 2005), this represents ~100–200 bullhead generations. It is therefore reasonable to assume that weirs represent a potentially relevant influence on both long- and short-term gene flow.

Microsatellite genotyping and variability: DNA was extracted from fin tissue using a standard “salting-out” method (BRUFORD *et al.* 1992). All individuals were genotyped at seven highly polymorphic microsatellite loci (ENGLBRECHT *et al.* 1999) using an ALFExpress automated sequencer (Amersham-Pharmacia). PCR conditions followed ENGLBRECHT *et al.* (1999), with minor modifications to annealing temperatures. Three or four internal size standards were used for each locus to permit reliable scoring of amplified fragments using FRAGMENT MANAGER 1.2 (Amersham-Pharmacia). Additionally, the same two individuals were run on all gels to facilitate cross-referencing among gels. Genotypes were checked for scoring errors that might be attributable to stutter products, large allele dropout, or the presence of null alleles, using the software MICRO-CHECKER v.2.2 (VAN OOSTERHOUT *et al.* 2004). Allele frequencies at each locus and for each sample were tested for conformity to Hardy–Weinberg equilibrium (as either an excess or deficit of homozygotes to maximize the power of tests) and for linkage disequilibrium among all combinations of loci using GENEPOP version 3.4 with 100 batches of 5000 iterations (RAYMOND and ROUSSET 1995). Provided that loci are unlinked, each can represent an independent test of a single hypothesis, such as a test of deviation from H–W (or population differentiation; see below), and, to combine probabilities, we applied the Bernoulli equation: $P = [N! / (N - K)! K!] \cdot \alpha^K (1 - \alpha)^{N - K}$, where N is the number of tests and K is the number below the designated type I error rate, α (MORAN 2003), which we set at 0.05. Following the terminology of KINNISSON *et al.* (2002), we refer to this as the “binomial likelihood method” (LM). Since the LM always requires repeatability of individually significant results across tests to yield an overall $P < 0.05$, it has the desirable property of not being influenced by single extremely low P -values in contrast with the commonly applied Fisher’s method. Genetic variability of samples was estimated as unbiased expected heterozygosity (H_e ; NEI 1978) and allelic richness (\hat{A})—the number of alleles, estimated by rarefaction, if all samples are equal to the smallest sample size—using FSTAT 2.9.3.2 (GOUDET 2001).

Equilibrium conditions: Two independent procedures were performed to test whether the set of populations from the River Rye had attained regional migration-drift equilibrium. The first procedure calculated the relative likelihood of a model of immigration–drift–equilibrium *vs.* a model of drift (without equilibrium) for the system as a whole, using a Markov chain Monte Carlo (MCMC) approach implemented in the program 2mod (CIOFI *et al.* 1999; software available at <http://www.rubic.rdg.ac.uk/~mab/software.html>). In 2mod the variable I indicates which of the alternative models is most likely across all populations for each iteration. The MCMC search was carried out twice for 10^5 iterations with the first 3×10^4 discarded as burn-in— I values proved to be wholly consistent in each run of 2mod. Our second procedure was based on the graphical isolation by distance (IBD) method of HUTCHISON and TEMPLETON (1999) that interprets the relationship between pairwise F_{ST} and distance in terms of outcomes expected with differing degrees of drift and migration. Specifically, HUTCHISON and TEMPLETON (1999) predicted that a strong IBD relationship throughout the sampled range, with increasing scatter as stochastic drift gradually

becomes more important than migration, is compatible with regional migration-drift equilibrium (type I relationship), whereas strong IBD at shorter distances that decays to a lack of detectable IBD at larger scales suggests a system exhibiting only local-scale equilibrium (type IV relationship; HUTCHISON and TEMPLETON 1999). Initially we used Mantel tests implemented by the software ZT (BONNET and VAN DE PEER 2002) with 100,000 randomizations to test the overall significance of the F_{ST} *vs.* distance relationship and saved residuals from a regression to test the relationship between scatter and distance, again using a Mantel test. However, since type IV relationships may be cryptic, we also split the data into distinct 15-km windows, which proved to be close to the minimum required to obtain enough data points for a reasonable signal of IBD, and within each window the F_{ST} *vs.* distance slope was analyzed separately. For the longer-distance pairwise comparisons in our study, F_{ST} might be impacted by mutation. Therefore, we also analyzed IBD relationships via $\log Nm$ *vs.* \log distance, since, at high levels of differentiation, Nm [calculated from $F_{ST} \approx 1 / (4Nm + 1)$] is less sensitive to mutation than F_{ST} (SLATKIN 1993).

Mutation-drift equilibrium was examined by testing for evidence of recent population bottlenecks [within the last ~0.2–4 N_e generations; LUIKART and CORNUET (1998) using the coalescent approach of CORNUET and LUIKART (1996) implemented in BOTTLENECK (PIRY *et al.* 1999). A recent simulation study (WILLIAMSON-NATESAN 2005) has shown that this test performed better than an alternative approach of GARZA and WILLIAMSON (2001) under conditions that are probable for bullhead populations studied using microsatellites: small prebottleneck diversity and a two-phase model of microsatellite evolution. The Wilcoxon signed-ranks procedure was used to test whether observed heterozygosity exceeded that expected at mutation-drift equilibrium under two more realistic models of microsatellite evolution: the stepwise mutation model (SMM) and a two-phase mutation model with 70% single-step mutations (TPM).

Population structure: To determine whether our samples might contain individuals that were first generation (F_0) immigrants from unsampled populations, we used the Bayesian assignment procedure of RANNALA and MOUNTAIN (1997), as implemented in GENECLASS 2 (PIRY *et al.* 2004). Since our specific task was to identify F_0 immigrants, we used PAETKAU *et al.*’s (2004) method to compute probabilities from 10,000 simulated genotypes. This creates a test distribution of simulated individuals by drawing haplotypes, rather than alleles, from the observed data and thus preserves the partial linkage disequilibrium present in genotypes that have immigrant ancestry, but are not F_0 immigrants (PAETKAU *et al.* 2004). WEIR and COCKERHAM’S (1984) method for estimating pairwise F_{ST} was calculated using GENEPOP 3.4. To compute an estimate of F_{ST} as a population-specific inbreeding coefficient (F), we used two different methods. F_{vit} was calculated using the method of VITALIS and COUVET (2001a) by the software ESTIM 1.2 (VITALIS and COUVET 2001b) and F_{2mod} by the software 2mod (CIOFI *et al.* 1999). Both methods estimate F from the probability that two genes share a common ancestor within a subpopulation, but while F_{vit} is estimated from analytical formulas, F_{2mod} is estimated from a coalescent model using a MCMC approach. We took the mean of F_{2mod} from two independent runs since estimates were highly correlated between runs (Pearson correlation, $R = 0.97$, $P < 0.001$). Probabilities of allelic (genic) homogeneity among populations were computed using the exact χ^2 test in GENEPOP 3.4 (using 100 batches of 5000 iterations), with the LM technique (see above) used to obtain a combined probability across loci.

Long-term effective population size and migration: Ratios of migration rate to mutation rate (M) and $\theta (= 4N_e\mu; N_e$ is

effective population size and μ is mutation rate) were calculated simultaneously for each pair of subpopulations using the coalescent-based maximum-likelihood (L_{MAX}) method implemented in MIGRATE 2.03 (BEERLI and FELSENSTEIN 1999, 2001). MIGRATE 2.03 assumes an n -island model at mutation–migration–drift equilibrium with values of M and θ constant over time. Since such constancy is generally unrealistic, M and θ are best viewed as long-term parameter estimates. MIGRATE uses a MCMC search algorithm to find the set of model parameters that maximizes the likelihood of observing the data. Due to the computational demands of MIGRATE, it proved necessary to limit the number of sample sites to 20 by excluding three locations from small tributaries of the River Seven (SuBe, NoBe, and HoBe) and one sample from the upper Hodge Beck (HoCT). These sites were chosen since we expected their levels of genetic exchange with the rest of the system to be relatively low, although we have also examined the possible effect of their omission (see below). The Brownian motion model was used as an approximation of the stepwise mutation model, and following initial trials, search criteria for the MCMC sampler were set to 20 short chains of 20,000 steps and 3 long chains of 200,000 steps. Searches were combined across all long chains (GEYER 1994) and a static heating scheme with four temperatures (1°, 1.5°, 3°, and 6°) was used (GEYER and THOMPSON 1994). MCMC sampling was carried out independently for each of the seven loci on separate computers and results were combined using the PERLscript program distributed with the MIGRATE package. The resulting combined file was used as the input for a separate run to obtain L_{MAX} values and confidence intervals across all loci using the “Fast” option. To obtain reliable parameter estimates, three subsequent runs were performed using the L_{MAX} estimates from the previous run as start values. Convergence was accepted if >95% of the confidence intervals for all parameters overlapped between runs. This criterion was achieved after the third run, at which point confidence intervals for all θ estimates, and for 96% of M estimates, overlapped with those from the previous run.

Estimates of gene flow between sample sites were calculated as Nm using four approaches. First, $Nm_{F_{\text{ST}}}$ estimates were calculated from pairwise F_{ST} -values via $F_{\text{ST}} \approx 1/(4Nm + 1)$ (WRIGHT 1951). Second, $Nm_{F_{2\text{mod}}}$ estimates were calculated, again using WRIGHT's (1951) equation, from the population-specific F -coefficients computed by 2mod (CIOFI *et al.* 1999). Third, $Nm_{F_{\text{it}}}$ estimates were calculated in the same way using F -values computed by Estim 1.2 (VITALIS and COUVET 2001b). Fourth, values of Nm_{MIG} were calculated for each population by multiplying MIGRATE's L_{MAX} estimates of M by $\theta/4$.

Recent migration rates: Patterns of very recent migration (over the last few generations) were investigated using the Bayesian admixture analysis (with the predefined populations option) in BAPS 3.1 (CORANDER *et al.* 2004). BAPS estimates the optimal posterior mode of the proportion of an individual's multilocus genotype that is represented by each population. Comparison of the likelihood of the modal solution with the likelihood of the individual being restricted to pure ancestry gives the posterior probability for the hypothesis of no recent immigrant ancestry. Using a critical probability level of 0.05, migration rates (m_{BAPS}) were calculated as the sum of genomic admixture proportions (either in total or from a specific population for pairwise comparisons) divided by the population sample size.

Testing the effect of unsampled populations: Although our sampling scheme was reasonably comprehensive, it is likely that unsuitable habitats such as deep, slow-flowing pools and lack of substrate impose geographic fragmentation on a scale smaller than that of our sampling grid. Consequently, it is possible that not all genetically distinct demes were sampled.

All of our estimators of total immigration as well as the pairwise estimates of Nm_{MIG} and m_{BAPS} are potentially sensitive to such unsampled subpopulations. We explored such potential sampling effects by comparing the results from the full data set with those obtained using a subset of 12 samples (supplemental Appendix 2 at <http://www.genetics.org/supplemental/>).

Hypothesis testing: First, we examined the interrelationships among several population genetic summary statistics, and also the geographical variable upstream size, using Spearman rank correlations—since a number of relationships were expected to be nonlinear—using SPSS v 11.5. Second, using ZT (BONNET and VAN DE PEER 2002), simple and partial Mantel tests were used to analyze the correlations among pairwise matrices of migration estimators and their relationships with pairwise geographic matrices, river distance, and the presence of weirs with 100,000 randomizations. Third, we examined the impact of weirs and distance on upstream and downstream migration using the directional estimates of long-term and recent immigration from MIGRATE and BAPS. Only samples from within the same tributary were used in this analysis because otherwise an unambiguous definition of upstream *vs.* downstream direction could not be made. Moreover, since Mantel tests must be applied to complete matrices, only a subset of sites that were connected in linear fashion could be used. Thus our analysis used one such linear subset, which comprised the River Seven sample sites (SeRF, SeTh, SeRo, SeHa, SeAs, SeSi, SeBB, SeBF, also including RyNB); other possible subsets contained too few sample sites that were not separated by weirs. Simple and partial Mantel tests were performed separately using upstream migration or downstream migration as dependent variables, with geographic distance and presence of weirs as independent variables. For pairwise comparisons involving these River Seven populations, the effect of weirs on the number of downstream-biased *vs.* upstream-biased migration estimates was tested directly using a contingency χ^2 test. Finally, the effect of distance and number of separating weirs on the change in allelic richness among neighboring River Seven sample sites was investigated via partial correlations (in SPSS v 11.5) with each predictor variable in turn held constant.

RESULTS

Exploratory data analysis: The seven microsatellite loci were highly polymorphic with between 9 and 44 alleles across all sample sites. MICRO-CHECKER revealed no evidence of scoring errors for any locus in any population. From a total of 504 tests, 22 exhibited significant evidence for linkage disequilibrium at a critical level of $P < 0.05$. This is slightly fewer than expected by chance as type I errors (25), and no tests were significant following sequential Bonferroni correction. Since populations can be regarded as independent tests for linkage disequilibrium between loci, it was possible to combine probabilities. No pairing of loci exhibited more than three test results below $P < 0.05$ (minimum LM $P > 0.08$). From 168 one-tailed tests, each for excess or deficiency of homozygotes, 6 and 4, respectively, exhibited $P < 0.05$. No more than one result of $P < 0.05$ for homozygote excess or deficit occurred in any single population across loci (LM $P > 0.25$), and no more than two of $P < 0.05$ for any locus across populations (LM $P > 0.20$). Thus, we detected no evidence for scoring errors,

linkage disequilibrium, or deviation from H-W equilibrium in our data.

Population structure: Using a critical P of 0.05 for first-generation (F_0) immigrant analysis, 15 individuals from 777 could not be assigned to any of the sampled locations; however, since the binomial probability of identifying only 15 immigrants by chance with a true type I error rate of 0.05 is $P < 0.001$, the power to detect F_0 immigrants appears to be low. To address this problem, we reset the critical P of the test to a new nominal level that yielded the number of migrants predicted by the upper binomial confidence limit (of 51, at a nominal critical P of 0.145). Moreover, to maintain a relatively liberal test, we made a Bonferroni adjustment of this critical P of 0.145 according to the sample size per population, rather than the overall sample size. At the Bonferroni-adjusted critical value of $P = 0.0045$ we identified only one migrant from outside the system—in the most downstream site, RyNB. Thus, while caveats regarding the power of the test must still be noted, it seems unlikely that migrants from outside the sampled sites will have had a substantial impact on our analyses.

There was marked genetic differentiation across all samples with a global F_{ST} of 0.27 (± 0.023 standard error) and pairwise F_{ST} estimates between 0.00 and 0.67 (supplemental Appendix 1 at <http://www.genetics.org/supplemental/>). Population differentiation was near universal among sampling locations with only 4 of 276 pairwise comparisons not significant (supplemental Appendix 1 at <http://www.genetics.org/supplemental/>). Differentiation was observed at very fine scales. For example, the comparison between SeAs and HoBe, which were located within the River Seven and a small tributary, respectively, yielded an F_{ST} estimate of 0.07 ($P < 0.001$) despite being only a few hundred meters apart.

Mutation-drift equilibrium: There were no significant tests for population bottlenecks under the most conservative mutation model, the SMM, although the test for SeRF approached significance ($P = 0.055$). However, four bottleneck tests were significant under the more realistic TPM (SeRF, $P < 0.05$; SeTh, $P < 0.05$; HoBe, $P < 0.05$; SuBe, $P < 0.01$), with the test for NoBe close to a significant level ($P = 0.055$). Therefore, while we did not detect evidence of deviation from mutation-drift equilibrium in most populations, five sample sites located upstream or in small tributaries show significant (or marginally significant) evidence of recent population decline.

Isolation by distance and migration-drift equilibrium: In every individual iteration of two runs (70,000/run), the indicator variable I (estimated by 2mod) suggested that a migration-drift equilibrium model was more likely than a nonequilibrium drift-only model. Moreover, within the river system as a whole there was a strong relationship between F_{ST} and distance (Mantel test, $R = 0.53$, $P < 0.001$) and increasing residual variation about the F_{ST} -distance regression with distance (Mantel test $R = 0.23$, $P < 0.05$). While such an IBD pattern appears

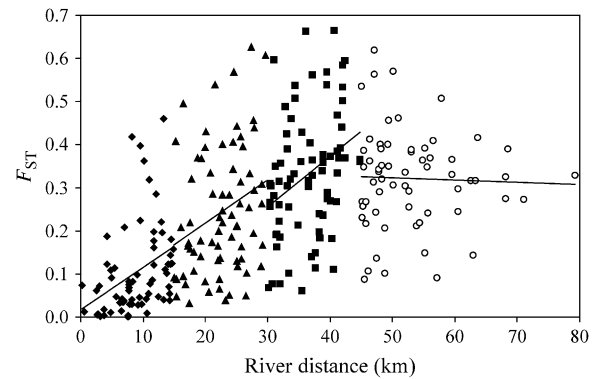


FIGURE 2.—Isolation by distance and approach to migration-drift equilibrium for F_{ST} . Separate regression lines are plotted for each 15-km window until 45 km and then from 45 to 80 km. The number of pairwise comparisons in each window are: 69, 72, 73, and 62, respectively, from left to right. The final window is shown with open symbols to highlight the change in relationship.

consistent with regional migration-drift equilibrium as predicted by HUTCHISON and TEMPLETON'S (1999) type I model, closer inspection of the F_{ST} -distance plot suggested that the relationship was not wholly linear with a possible change in slope between 40 and 50 km. Indeed, when split into four nonoverlapping distance windows (three of 15 km and a final one of 35 km; Figure 2), we identified strong and very similar patterns of IBD to 45 km, but no F_{ST} -distance relationship in the final window (45–80 km). Since data points arising from a pairwise matrix are not independent, the significance of slopes was not tested directly, but instead we divided the data into four groups of six sample sites on the basis of their average separation from all other sites (Table 1). Consistent with the pattern of slopes in Figure 2, the IBD relationship was significantly positive in the first three distance groupings, but not significant in the distance window with the highest average pairwise distance (Table 1). Thus, the river system appeared to show local migration-drift equilibrium to a scale of ~ 45 km but had not attained equilibrium at longer distances. Since the maximum distance between sampling points within any individual river is 45.5 km (RaBe and RyNB; see Figure 1), evidence for nonequilibrium conditions arose primarily from comparisons among rivers, particularly among upstream locations.

Evidence of past colonizations: The shape of the residuals about the regression line in the final window of Figure 2 is of interest because of an apparent decline with distance. We examined this nonequilibrium region in more detail via the framework of SLATKIN (1993), who showed that the slope of $\log Nm_{F_{ST}}$ vs. \log distance in a one-dimensional stepping-stone model can change progressively from negative to zero or positive, with decreasing residual scatter, depending on time since colonization. Two extreme distance windows were partitioned: (1) short-distance comparisons (maximum pairwise

TABLE 1
Distance characteristics (mean, min, max) of the windows partitioned to analyze variation in the correlation (Pearson's R) between F_{ST} and distance

Window	Mean	Min	Max	R	P	Sites
1	9.6	2.7	17.5	0.75	0.031	RyNB SeBB SeBF RyWN DoSH SeSi
2	19.6	0.2	37.2	0.77	0.001	DoKM SeAs HoBe HoHC SuBe SeHa
3	37.5	0.6	47.1	0.42	0.035	DoLo RyDu BogG HOCT SeRo NoBe
4	46.5	2.6	79.3	0.10	0.346	SeTh RyCr DoHF SeRF RyHa RaBe

Probabilities are from Mantel tests.

separation of 15 km and a mean of 9 km) comprising the five most downstream sites in the system and (2) long-distance comparisons (minimum pairwise separation of 30 km and a mean of 51 km) comprising the most upstream site in each major river. There was a striking contrast in slopes between the distance windows, with a very strong negative correlation between $\log Nm_{F_{ST}}$ and \log distance for short-distance comparisons (and no significant trend in residuals) yet a similarly strong positive correlation for long-distance comparisons, coupled with significantly decreasing residual variation (Figure 3). Therefore, short-distance comparisons among downstream sites show an IBD relationship expected at equilibrium in a stepping-stone population model. Long-distance comparisons among extreme upstream sites, on the other hand, apparently carry a strong historical colonization signal, consistent with SLATKIN's (1993) predictions for a recent expansion scenario in a linear stepping-stone model.

Effects of unsampled ghost populations on parameter estimates: Full details of our investigation, which compared results from four different analytical methods with those obtained using a subset of 12 sample sites, are given in supplemental Appendix 2 at <http://www.genetics.org/supplemental/>. Measures of total immi-

gration (Nm_{MIG} and m_{BAPS}), inbreeding coefficients (F_{vit} and F_{2mod}), and θ all showed a very strong correlation between full and reduced data sets, as did pairwise m_{BAPS} and, to a somewhat lesser extent, Nm_{MIG} . However, while the F -coefficients and θ were not consistently higher or lower in the reduced data set compared to the full data set, total Nm_{MIG} and m_{BAPS} showed a significant difference between data sets with the former $\sim 40\%$ higher, and the latter 40% lower in the reduced data set. While pairwise m_{BAPS} values were not consistently higher or lower in the reduced data set compared to the full data set, pairwise estimates of Nm_{MIG} differed by 176% in the reduced data set, with 89% of values higher than those in the full data set. Thus for our study system, unsampled ghost populations appear to have little potential to affect estimates of F -coefficients, θ , and pairwise m_{BAPS} , and a moderate, although directional, impact on total m_{BAPS} and Nm_{MIG} . However, since pairwise Nm_{MIG} values were on average substantially higher in the reduced data set, both ghost populations and the four sites intentionally omitted from analysis (see MATERIALS AND METHODS) could have exerted a nontrivial quantitative impact on results.

Genetic diversity, population isolation, and migration asymmetry: Genetic diversity, whether measured as \hat{A} or H_e , varied about fivefold across the sample sites, as did estimates of θ (Table 2). Measures of genetic diversity (\hat{A} and H_e) and θ were strongly interrelated, and each was significantly correlated with upstream size (Table 3), indicating that upstream locations displayed lower genetic diversity and N_e than their downstream counterparts. The inbreeding coefficients F_{2mod} and F_{vit} showed strong inverse correlations with θ , \hat{A} , H_e , and upstream size (Table 3), demonstrating that smaller, less genetically diverse upstream populations were also more genetically isolated. A mechanism potentially contributing to these patterns was evident from both immigrant gene flow, Nm_{MIG} , and recent immigration rates, m_{BAPS} , both of which were reduced to more upstream populations (Tables 2 and 3). Clear evidence for systematic source-sink population structure was provided by patterns of immigration–emigration asymmetry ($Im-Em$), a measure of the degree to which a population is a donor or a recipient of migrants. $Im-Em$ increased strongly with upstream size (Table 3); *i.e.*,

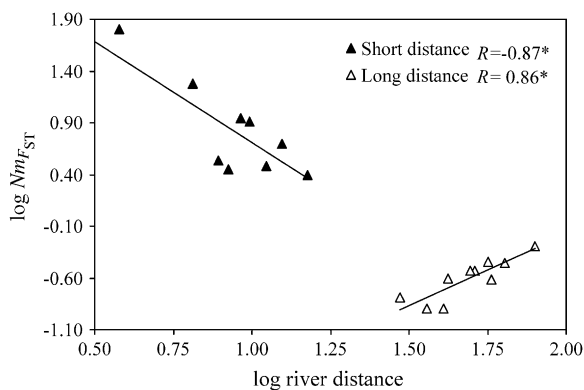


FIGURE 3.—Plot of $\log Nm_{F_{ST}}$ against \log river distance. Pairwise comparisons between two groups of sample sites comprise the five most downstream locations (yielding short-distance comparisons) and the five most upstream locations in separate rivers (yielding long-distance comparisons). R is the correlation coefficient; *Mantel tests, $P < 0.05$.

TABLE 2
Characteristics of each sample site

Site	<i>n</i>	Upstream size	H_e	\hat{A}	θ	F_{2mod}	F_{vit}	Nm_{MIG}	m_{BAPS}
RaBe	33	1.50	0.54	4.55	0.32	0.24	0.23	4.08	0.03
RyHa	35	4.00	0.50	6.15	0.46	0.11	0.29	4.30	0.06
RyCr	31	24.19	0.59	7.25	0.48	0.09	0.16	5.83	0.17
RyDu	35	31.09	0.60	7.48	0.51	0.07	0.14	8.05	0.25
RyWN	37	48.52	0.63	9.19	0.65	0.05	0.11	6.03	0.15
BogG	35	1.00	0.28	2.59	0.22	0.67	0.61	1.36	0
DoHF	40	2.50	0.38	2.86	0.27	0.38	0.46	2.07	0.04
DoLo	35	11.70	0.44	3.87	0.29	0.26	0.38	2.92	0.15
DoKM	40	19.91	0.54	5.18	0.41	0.18	0.23	6.03	0.21
HoCT	34	2.00	0.14	1.65	NA	0.52	0.80	NA	0
HoHC	34	7.40	0.19	2.20	0.20	0.41	0.73	1.13	0.06
DoSH	35	40.29	0.64	7.76	0.53	0.07	0.09	7.73	0.24
SeRF	18	0.50	0.51	2.86	0.17	0.42	0.28	2.68	0.04
SeTh	26	3.10	0.56	3.62	0.22	0.28	0.21	3.66	0.27
SeRo	38	6.20	0.52	3.44	0.32	0.29	0.26	6.24	0.20
NoBe	23	0.50	0.54	3.36	NA	0.30	0.23	NA	0.22
SeHa	36	11.30	0.52	3.70	0.33	0.26	0.27	5.16	0.37
SeAs	36	14.10	0.55	3.80	0.34	0.38	0.22	4.96	0.11
HoBe	35	0.50	0.57	3.95	NA	0.22	0.19	NA	0.09
SuBe	23	4.50	0.65	4.18	NA	0.28	0.07	NA	0.09
SeSi	33	29.20	0.63	4.84	0.45	0.15	0.11	6.30	0.32
SeBF	34	34.19	0.68	8.20	0.69	0.05	0.03	11.17	0.30
SeBB	22	36.89	0.73	8.65	0.53	0.04	-0.03	9.27	0.40
RyNB	29	158.86	0.74	9.66	0.85	0.02	-0.05	14.85	0.47

Values are: sample size (*n*), upstream size and estimates of genetic diversity (H_e , \hat{A}), relative effective population size (θ), population inbreeding coefficients (F_{2mod} , F_{vit}), total immigrant gene flow (Nm_{MIG}), and total recent immigration rate (m_{BAPS}).

upstream populations tended to be donors and downstream populations to be recipients (Figure 4).

Comparison between estimators of migration: Gene flow estimated from F_{2mod} showed a strong correlation with gene flow estimated as Nm_{MIG} (Figure 5). While the range of values was very similar, the relationship was nonlinear, with $Nm_{F_{2mod}}$ much lower than Nm_{MIG} values for populations with low Nm , but correspondingly more

closely related at higher levels of Nm (Figure 5). $Nm_{F_{vit}}$ values showed a similar pattern, although Nm could not be estimated from the negative estimates of F_{vit} for two of the populations. Although highly significant, the correlation between pairwise matrices of $Nm_{F_{ST}}$ and Nm_{MIG} was moderate (Mantel test, $R = 0.33$, $P < 0.001$) and weak between $Nm_{F_{ST}}$ and m_{BAPS} ($R = 0.19$, $P < 0.05$). However, the relationship between pairwise matrices of

TABLE 3

Spearman rank correlation coefficients (below diagonal; probabilities above) between estimators of genetic diversity, inbreeding coefficients, and migration for each sample site, and the environmental predictor variable upstream size

	Upstream size	H_e	\hat{A}	θ	F_{2mod}	F_{vit}	Nm_{MIG}	m_{BAPS}	$Im-Em_{MIG}$	$Im-Em_{BAPS}$
Upstream size		<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001
H_e	0.658		<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001
\hat{A}	0.787	0.832		<0.001	<0.001	<0.001	<0.001	<0.01	<0.001	<0.001
θ	0.890	0.842	0.957		<0.001	<0.001	<0.001	<0.01	0.000	<0.001
F_{2mod}	-0.773	-0.795	-0.971	-0.943		<0.001	<0.001	<0.001	<0.001	<0.001
F_{vit}	-0.655	-0.999	-0.827	-0.847	0.789		<0.001	<0.001	<0.001	<0.001
Nm_{MIG}	0.831	0.892	0.837	0.887	-0.852	-0.897		<0.001	<0.001	<0.001
m_{BAPS}	0.676	0.706	0.606	0.626	-0.680	-0.709	0.806		<0.001	<0.001
$Im-Em_{MIG}$	0.792	0.881	0.792	0.785	-0.821	-0.886	0.914	0.807		<0.001
$Im-Em_{BAPS}$	0.743	0.753	0.691	0.764	-0.750	-0.755	0.859	0.863	0.850	

Symbols are as in Table 1, except $Im-Em_{MIG}$ is mean immigrant-emigrant gene flow and $Im-Em_{BAPS}$ is mean recent immigration-emigration rates (see Figure 4).

$n = 20$ for parameters calculated by MIGRATE (see MATERIALS AND METHODS) and $n = 24$ for all others.

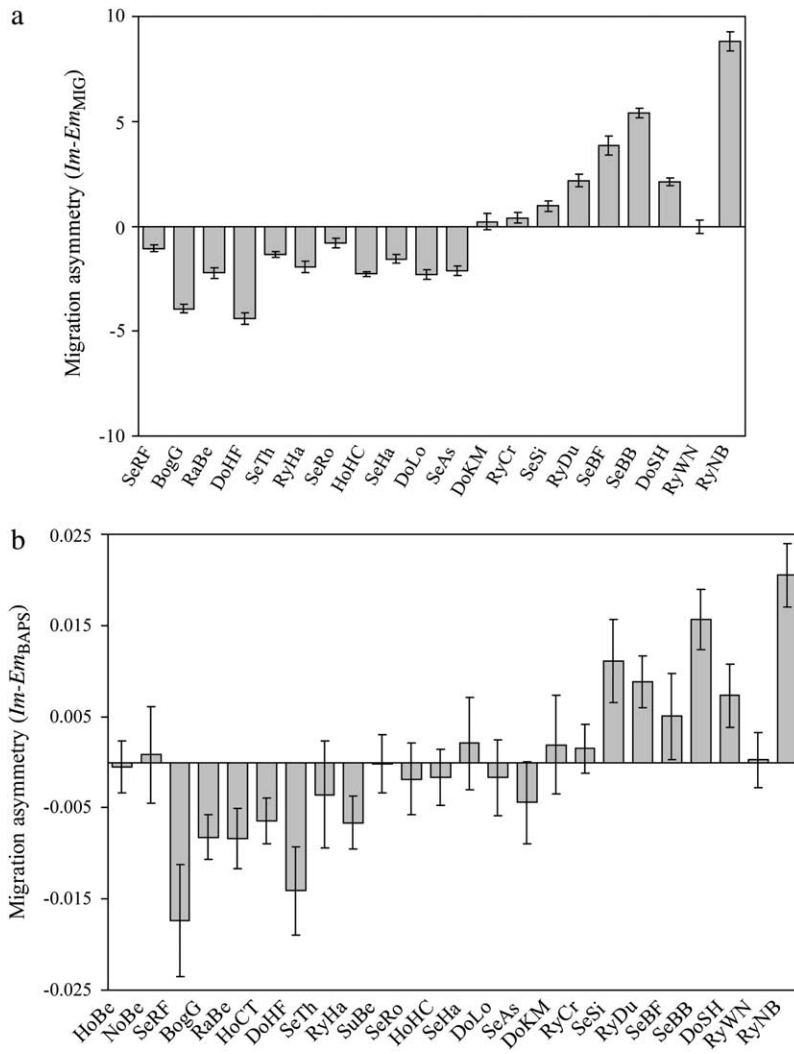


FIGURE 4.—Immigration–emigration asymmetry in (a) long-term gene flow (Nm_{MIG}) and (b) recent migration rate (m_{BAPS}). Bars show the mean of differences between immigrant and emigrant estimates per population, ± 1 standard error. Samples sites are ordered (left to right) by upstream size, such that the more downstream locations are to the right (see Figure 1 for locations).

F_{ST} and Nm_{MIG} or m_{BAPS} was much stronger (Mantel tests, $R = -0.76$, $P < 0.001$ and $R = -0.53$, $P < 0.001$, respectively), suggesting that the lower congruence of $Nm_{F_{ST}}$ and Nm_{MIG} resulted from the translation of pairwise F_{ST} values to Nm . All pairwise estimates of $Nm_{F_{ST}}$ were higher than corresponding Nm_{MIG} estimates (supplemental Appendix 1 at <http://www.genetics.org/supplemental/>), as was global $Nm_{F_{ST}}$ (0.68) by comparison to the overall average Nm_{MIG} (0.28). Pairwise directional estimates of scaled migration rates M_{MIG} were quite well correlated with m_{BAPS} for both immigration ($R = 0.49$, $P < 0.001$, Mantel test) and emigration ($R = 0.54$, $P < 0.001$) and particularly so for the average of immigration and emigration for each pair of populations (Mantel test, $R = 0.71$, $P < 0.001$). However, Figure 6 suggests that m_{BAPS} declined much more sharply with distance than M_{MIG} .

Flow direction, weirs, and distance as determinants of migration: To examine whether distance and weirs affected migration in the river system as a whole, we applied simple and partial Mantel tests to the pairwise matrices of Nm_{MIG} , m_{BAPS} , and F_{ST} . Distance and weirs

showed significant correlations with all metrics in both simple and partial Mantel tests (Table 4a). The preceding analysis had the advantage of including all or most sample sites, but had limited power to estimate the relative impacts of weirs and pairwise distance on gene flow because of relatively few sites that were not separated by weirs. Moreover, the action of flow direction on migration might be an important determinant of the relationships between upstream size and population genetic characteristics, but could be assessed unambiguously only among sites within the same river. Therefore, we compared estimates of upstream *vs.* downstream Nm_{MIG} for sites within the linearly arranged chain of populations from the River Seven and Ry/NB. In addition, with a maximum pairwise distance of 31.5 km (mean = 13.8 km) sample sites along the River Seven were likely to be near local migration-drift equilibrium. Downstream-biased migration was significantly more frequent than expected by chance for both Nm_{MIG} ($\chi^2 = 13.4$, $P < 0.001$) and m_{BAPS} ($\chi^2 = 6.8$, $P < 0.01$). However, as Figure 7a illustrates, detection of a downstream bias in Nm_{MIG} was dependent on the presence of a

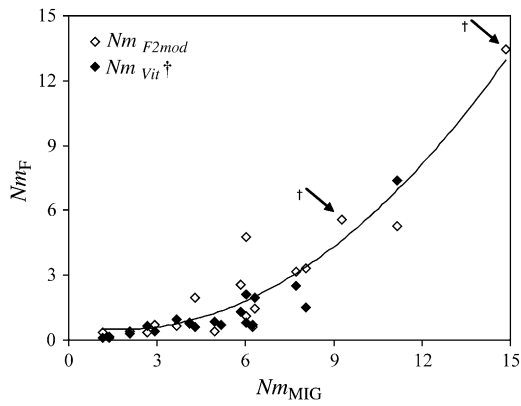


FIGURE 5.—Comparison of Nm calculated from inbreeding coefficient (F) estimates for each population (F_{2mod} and F_{vit}) via $F \approx 1/(1 + 4Nm)$ with total estimated immigration from MIGRATE (Nm_{MIG}). A quadratic line is fitted to the $Nm_{F_{2mod}}$ data ($R^2 = 0.91$), which provided a better fit than a linear function ($R^2 = 0.78$). †x-axis positions of missing data points from $Nm_{F_{vit}}$, where F_{vit} estimates were negative and thus Nm was inestimable.

weir between sites (Fisher's exact test, $P < 0.01$). Figure 7b suggests a similar effect of weirs on asymmetry for m_{BAPS} , and although this was marginally nonsignificant (Fisher's exact test, $0.05 < P < 0.10$), fewer data were available because of some cases of zero m_{BAPS} estimates in both flow directions. Weirs reduced upstream migration to a greater extent than distance, for both Nm_{MIG} and m_{BAPS} , although distance remained independently significant, or almost so, for both estimators (Table 4b). By contrast, distance had little independent effect on downstream migration for Nm_{MIG} , and the negative impact of weirs, although significant, was considerably less than for upstream migration (Table 4b). For m_{BAPS} , on the other hand, the effect of distance on downstream migration was much greater than that of weirs (Table 4b).

A possible consequence of these relationships on genetic diversity is illustrated in Figure 8. Change in \hat{A} between neighboring sample sites in the River Seven was significantly correlated with the number of weirs (partial correlation with neighbor distance held constant, $R = 0.85$, $P < 0.05$), but showed no independent correlation with neighbor distance (partial correlation with number of weirs between neighbors held constant, $R = 0.09$, NS).

DISCUSSION

Distribution of genetic diversity: The consistently high, and significant, levels of genetic differentiation even at a small geographical scale (>0.5 km) indicated that bullhead populations of the River Rye catchment are highly fragmented despite a near-continuous distribution of suitable habitat across the study area. This is consistent with field data that indicate a low individual dispersal capacity of bullheads (DOWNHOWER *et al.*

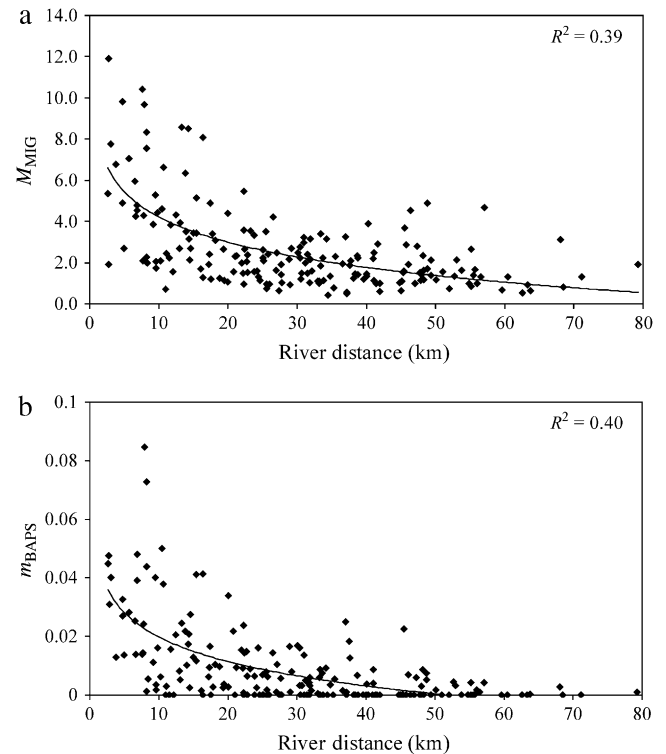


FIGURE 6.—Variation in pairwise migration rates with increasing distance between locations. (a) M_{MIG} is the estimated long-term migration rate/mutation rate; a logarithmic curve (R^2 shown) provided a better fit than a linear regression ($R^2 = 0.27$). (b) m_{BAPS} is the estimated recent migration rate; again, a linear regression provided a poorer fit ($R^2 = 0.29$) than the fitted logarithmic curve.

1990; KNAEPKENS *et al.* 2004) and marked genetic (allozyme) differentiation among neighboring river systems (HÄNFLING and BRANDL 1998a). Furthermore, there was a striking level of correlation among nearly all measures of genetic diversity and migration/gene flow and also with upstream size—a measure of position within the river system. Therefore, in relative terms at least, many genetic characteristics of bullhead populations within the river system could be predicted from consideration of a small number of measures or, in the absence of specific genetic data, by consideration of the upstream location of a population. Upstream samples exhibited significantly lower levels of genetic diversity and lower effective population sizes and receive fewer immigrants than their downstream counterparts. A strong relationship between upstream size and allelic richness was previously described when comparing bullhead populations from distinct river systems across Europe, including one site (RyWN) from the present data set ($R = 0.76$, $P < 0.001$; HÄNFLING *et al.* 2002). With a near identical correlation, this study strongly suggests that the result in HÄNFLING *et al.* (2002) was not simply attributable to large-scale variation among long-diverged populations, which otherwise might be most parsimoniously explained by historical processes.

TABLE 4

Pearson correlations between pairwise distance and the presence of a weir separating sample sites and (a) pairwise estimates of migration or population differentiation for all sample sites and (b) directional migration estimates in the River Seven

	Simple Mantel test		Partial Mantel test	
	Distance	Weir	Distance	Weir
a.				
Nm_{MIG}	-0.56***	-0.51***	-0.46***	-0.39***
m_{BAPS}	-0.54***	-0.40***	-0.46***	-0.25**
F_{ST}	0.53***	0.35***	0.46***	0.18*
b.				
Nm_{MIG}				
Upstream	-0.60***	-0.79***	-0.37*	-0.70***
Downstream	-0.31*	-0.48**	-0.09	-0.39*
m_{BAPS}				
Upstream	-0.47***	-0.58**	-0.26 ^a	-0.44**
Downstream	-0.40*	-0.16	-0.37*	0.05

*** $P < 0.001$; ** $P < 0.01$; * $P < 0.05$.

^a $0.05 < P < 0.10$.

Scale of equilibrium: Even at the small geographic scale of sampling in this study, the strong and predictable pattern in the distribution of genetic diversity could be generated either through historical processes (*e.g.*, past upstream colonization) or by ongoing ecological processes (*e.g.*, downstream-biased gene flow). Genetic signals of past colonization events are expected to disappear over time as populations approach migration-drift equilibrium. Two independent tests—the strength of isolation by distance, the comparison of drift *vs.* migration + drift models (in the software 2mod), and additionally, the strong correlation between long-term and contemporary migration—each suggested that bull-head populations from the River Rye catchment are to a large extent in migration-drift equilibrium. Nevertheless, close examination of the F_{ST} -distance relationship revealed a pattern consistent with the HUTCHISON and TEMPLETON (1999) type IV graphical model, suggesting that while much of the system is in migration-drift equilibrium, long-distance comparisons involving upstream sites will reflect historical, rather than contemporary, gene flow. This finding highlights the importance of careful examination of evidence for migration-drift equilibrium throughout different parts of a natural study system. Acceptance of more general evidence such as that of 2mod, which tests only extreme scenarios (equilibrium or not), can result in heterogeneity of equilibrium conditions being overlooked. Indeed, where a stepping-stone population model seems most applicable, it is perhaps wise to consider lack of migration-drift equilibrium as the null hypothesis, at least when considering sites at the edges of a distribution, because the

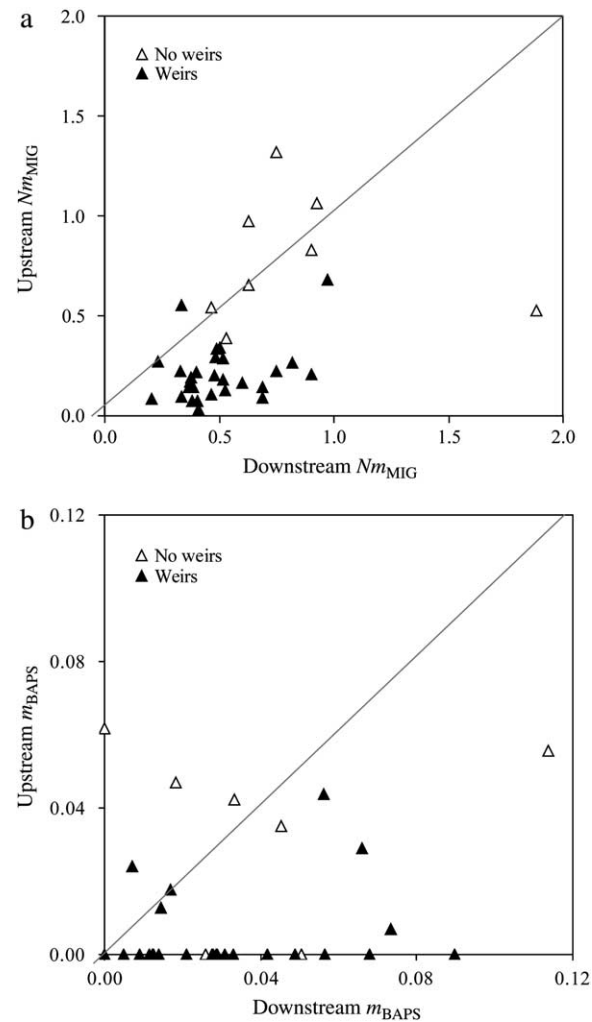


FIGURE 7.—Upstream *vs.* downstream (a) long-term gene flow and (b) recent migration in a linear chain of nine samples sites from the River Seven (shown in Figure 8). Pairwise comparisons separated by at least one intervening weir are coded by open symbols; those without an intervening weir by solid symbols. The diagonal line in each shows a 1:1 upstream:downstream relationship.

time to approach equilibrium will often be extremely long (SLATKIN 1993; EFREMOV 2004)

Evidence for extinction-recolonization dynamics: Bottleneck analysis indicated that at least four locations have experienced a recent population decline. All were in sites with low upstream size, either in small tributaries or in the upper part of the catchment, suggesting that reductions in population sizes might occur primarily at the upstream margins of the river system. While the low power of all bottleneck tests (WILLIAMSON-NATESAN 2005) means that we cannot rule out the possibility that downstream sites have also experienced recent declines, our *a priori* expectation was that isolated upstream sites would be more prone to such events. Such bottlenecks might reflect either *in situ* decline of preexisting populations or founder events arising from colonizations. Evidence for colonizations came from comparison of

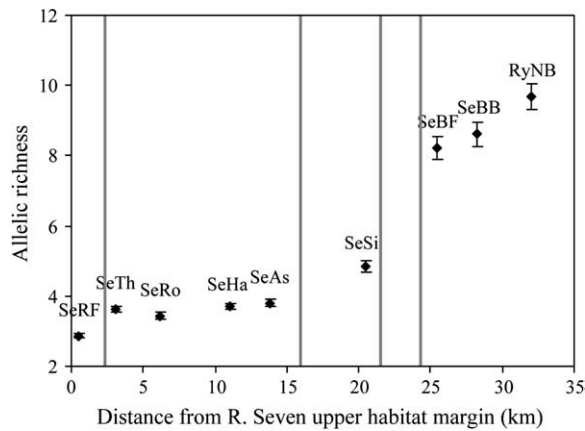


FIGURE 8.—Change in allelic richness (\hat{A}) in relation to distance between neighboring populations (measured from the top of the River Seven) and intervening weirs (vertical lines) in a linear chain of sample sites from the River Seven. Points show the mean \hat{A} across microsatellite loci with standard errors obtained by jackknifing over loci.

correlations between $\log Nm_{F_{ST}}$ and \log distance for extreme upstream and extreme downstream locations. As predicted by SLATKIN'S (1993) coalescent simulations, the negative relationship expected at equilibrium was strong among downstream locations, but reversed (*i.e.*, strongly positive) among upstream locations, which also, as predicted, showed declining residual variation with distance. SLATKIN'S (1993) simulations suggest that such a positive slope will persist for only a very short time in a one-dimensional stepping-stone model (<10 generations) before decaying to a more temporally stable IBD state of approximately zero slope. Although the timescale of Slatkin's predictions might be affected by deviations from his stepping-stone range expansion model, such a strong colonization signal is likely of recent origin and thus more likely reflects recolonizations of previously occupied areas, rather than a novel range expansion.

It might seem puzzling that only one (SeRF) of the five populations representing the upstream extremities of our sampling showed evidence for a bottleneck, and none for linkage disequilibrium, which might also be expected. However, extreme upstream locations were genetically impoverished, even to the extent that some loci were monomorphic. Consequently, the power to detect both bottlenecks and linkage disequilibrium (see, for example, ZAYKIN *et al.* 1995 for the latter) is likely to have been particularly low for upstream locations. Alternatively, if Slatkin's predicted timescale is applicable, post-bottleneck heterozygosity may have had insufficient time to recover for a bottleneck signal to be detected ($\approx 0.2 N_e$ generations; LUIKART and CORNUET 1998).

Natural vs. anthropogenic determinants of source-sink structure: Relative N_e (*i.e.*, θ) and immigration estimates showed a strong correlation with upstream

size. Gene flow was strongly biased downstream with lower N_e upstream populations serving as net producers of migrants and higher N_e downstream populations as net recipients. Such asymmetric gene flow could be the result of the passive downstream drift created by the essentially unidirectional water movement in riverine environments. Therefore, one might expect such a pattern to be a natural consequence of living in a river habitat. All else being equal, this should prevent long-term persistence of riverine populations, a phenomenon known as the "drift paradox" (MÜLLER 1954) that has puzzled ecologists for >50 years. Most solutions proposed to solve the drift paradox indicate that either active or passive compensatory upstream migration is required to prevent a species boundary from shifting downstream (HERSHEY *et al.* 1993; SPEIRS and GURNEY 2001; HUMPHRIES and RUXTON 2002; PACHEPSKY *et al.* 2005). Our analysis of data from the River Seven showed that upstream migration was much more strongly impacted by the presence of weirs than downstream migration. Moreover, stepwise changes in allelic richness among these neighboring populations was unrelated to distance among sites but strongly correlated with the number of weirs. Although the populations of the River Seven are within a distance window that suggests that they could be near migration-drift equilibrium, the good agreement between results for long-term gene flow, Nm_{MIC} (which assume equilibrium), and for recent migration rates, m_{BAPS} (which do not), adds considerable strength to our conclusions. The only disagreement between estimators concerned the relative effects of distance and weirs on downstream migration, which may result from the different timescale that each estimator can detect (see below).

Our results suggest that bullhead populations may be able to at least partly compensate for downstream drift by active upstream migration, provided this is not prohibited by barriers. In turn, this suggests that anthropogenic obstructions are a major source of habitat fragmentation and enhance the isolation and low genetic diversity of upstream bullhead populations. Such anthropogenic impacts might threaten population persistence in at least three ways. First, fragmentation might increase the risk of stochastic population extinction, which is particularly high in hierarchical dendritic habitats such as rivers (FAGAN 2002; MORITA and YOKOTA 2002). Second, a "washout" effect caused by prevention of the upstream movement necessary to overcome passive drift could lead to a gradual shift of distribution boundaries downstream (SPEIRS and GURNEY 2001). Third, inbreeding depression caused by genetic isolation might lower survival probabilities of upstream individuals and further diminish population sizes (SACCHERI *et al.* 1998; BROOK *et al.* 2002). Previous work suggests that inbreeding depression occurs in isolated Belgian bullhead populations (KNAEPKENS *et al.* 2002b), and we are currently investigating whether the

same may apply in some of the upstream populations of the River Rye system.

Recent results from the grayling *Thymallus thymallus* (MELDGAARD *et al.* 2003) and the cutthroat trout *Oncorhynchus clarki clarki* (WOFFORD *et al.* 2005) have shown, respectively, a correlation between numbers of weirs and F_{ST} and between numbers of isolating barriers (some anthropogenic) and genetic diversity (H_e , \hat{A}). Although neither article investigated migration as directly as this study, or the interaction between weirs and flow direction, the general comparability of results from these highly mobile, fecund salmonids and a sedentary, low-fecundity sculpin (*C. gobio*) suggests that anthropogenic barriers could have substantial negative impacts on river fish in general.

Congruence among estimators of migration:

The population structure of bullheads in the River Rye system is characterized by distance-limited dispersal with directional barriers to migration, systematic asymmetry in Nm and N_e , deviation from equilibrium, and perhaps also extinction–recolonization dynamics. Consequently, our study system bears little resemblance to WRIGHT's (1951) infinite island model upon which the $F_{ST}:Nm$ relationship is based. Although inference of gene flow from F_{ST} seems quite robust to a number of deviations from the island model, including moderate numbers of subpopulations and distance-limited dispersal (reviewed by (NEIGEL 2002), systematic asymmetry in N_e , m , or Nm can be problematic for estimation of $Nm_{F_{ST}}$ (WHITLOCK and McCAULEY 1999; WAKELEY 2001; WILKINSON-HERBOTS and ETTRIDGE 2004). Moreover, F_{ST} can be biased downward by mutation (ROUSSET 1996), which can be considered negligible only if $m \gg \mu$ (where m and μ are mutation and migration rates, respectively). Even at the fine geographic scale of sampling in our study, migration between populations was often very limited and MIGRATE's estimates of M ($= m/\mu$) ranged from 0.2 to 13.2 (supplemental Appendix 1 at <http://www.genetics.org/supplemental/>). Such estimates are plausible when highly polymorphic microsatellites are used to screen individuals from a highly differentiated system and suggest that mutation might have impacted some of our estimates of F_{ST} . By contrast, MIGRATE incorporates mutation, albeit using a simple model of microsatellite evolution and estimates relative m in both directions among populations and also relative N_e for each population. Consequently, since MIGRATE is less constrained by assumptions than $Nm_{F_{ST}}$, MIGRATE parameter estimates are often considered likely to be more accurate, at least provided reasonably convergent estimates can be obtained (as was the case for our study). Unfortunately, because of its computational demands, tests of MIGRATE have been limited to date, and although BEERLI and FELSENSTEIN (1999) reported good performance in their initial article describing the model, recent tests using simulated mtDNA sequence data by ABDO *et al.* (2004)

showed reasonable accuracy in the estimation of relative N_e , but often substantial inaccuracy in estimates of M . The generality of ABDO *et al.*'s (2004) results is not clear, especially for microsatellite data, but they highlight the importance of comparing parameter estimates from MIGRATE with those obtained from other methods, since in natural systems true parameter values are unknown.

In contrast to the strong correlations between different population-specific estimates of total immigrant Nm , pairwise estimates of Nm_{MIG} (average of immigration and emigration) were only moderately well correlated with pairwise $Nm_{F_{ST}}$ (calculated from Weir and Cockerham's estimator of F_{ST}), and all pairwise $Nm_{F_{ST}}$ estimates were lower than corresponding Nm_{MIG} values. This contrasted with the population-specific Nm estimates from F_{2mod} and F_{vit} , which were always lower than the total Nm_{MIG} , and might reflect a much greater sensitivity to mutation for F_{ST} calculated from pairwise allele-frequency variation than F_{ST} calculated as an inbreeding coefficient (NEIGEL 2002). We are not aware of any previous comparisons of gene flow estimates from F_{2mod} or F_{vit} with those from MIGRATE, but the relationships proved interesting since estimates shared a similar range but became progressively less congruent as estimated Nm decreased (see Figure 5). One possible explanation for this pattern is that, by contrast to MIGRATE, neither method for estimating F allows for asymmetry in N_e , which can inflate F (WAKELEY 2001) and consequently would deflate Nm . Such underestimation will become more pronounced as populations show a higher drift:immigration balance and could be important in our system where N_e and immigration appear to co-vary. Alternatively, while it is possible that Nm_{MIG} values were overestimated—either because of ghost populations (see below) or because of nonequilibrium signals of past migration—it is not clear why they should have become progressively less so relative to those from the F -coefficients, which also rely on equilibrium.

In our study system, and probably many others where the scale of equilibrium was tested less explicitly, the assumption of migration-drift equilibrium was not wholly fulfilled. Given this limitation of equilibrium-based methods such as MIGRATE, comparisons with nonequilibrium estimators of contemporary migration, such as BAPS, may be informative. Migration estimates from MIGRATE and BAPS were well correlated whether estimated as totals per population or as either directional or nondirectional (averaged) pairwise estimates. The estimators diverged somewhat when patterns of variation across pairwise distance among populations were compared, with m_{BAPS} declining more sharply with distance than M_{MIG} . This difference might be largely attributable to a persistent signal of past colonization in longer distance Nm_{MIG} estimates, but could also be partly due to the greater capability of MIGRATE than BAPS to detect indirect migration events in a population structure resembling a stepping-stone model.

How important were ghost populations? Few studies of natural systems will be able to sample all subpopulations that might exchange migrants. While the estimates produced by traditional measures of pairwise F_{ST} and population genetic diversity are independent of the sampling scheme, this cannot be assumed to be true for modern methods that estimate all parameters of interest simultaneously. Yet there have been few tests to date of the impact of unsampled ghost populations (as defined by SLATKIN 2005). Recently, both BEERLI (2004) and SLATKIN (2005) have tested the sensitivity of MIGRATE parameter estimates to unsampled populations but came to different conclusions, with Beerli suggesting a limited effect, but Slatkin suggesting that impacts were difficult to predict, but could be large with substantial immigration from ghost populations.

Our results from evaluating potential impacts of ghost populations illustrate a rather extreme example since we removed up to half of the sample sites. Indeed, the analysis used to detect first-generation immigrants suggested that few immigrants could not be assigned to one of the populations that we had sampled. Consequently, ghost populations probably had a limited effect on most of our results since they most likely represented “intermediate” subpopulations in a linear stepping-stone model. Nevertheless the apparent sensitivity of the different estimators was interesting. Estimators of F -coefficients were not impacted strongly by use of a reduced data set. This was anticipated for the method of VITALIS and COUVET (2001a,b), but was far less certain for that implemented by 2mod (CIOFI *et al.* 1999). The effects on MIGRATE’s estimates of θ and Nm differed, with θ well correlated and not directionally biased between the data sets, but total Nm_{MIG} and particularly pairwise Nm_{MIG} estimates showing a significant upward bias. Since, even for our full MIGRATE data set, we omitted 4 of the 24 populations to reduce run times, it is likely that our estimates of migration parameters were upward biased to some extent. Use of reduced data sets for MIGRATE is quite commonplace in the literature because of computational demands, but our analysis suggests that resulting parameter estimates may well be biased. Sampling effects on BAPS were quite different from those for MIGRATE, with a significant downward bias to total immigration rate estimates but virtually no effect on pairwise estimates. This rather surprising result implies that immigrant genotypes, or portions thereof, were not reallocated to the “next best” population if their most probable source is omitted, although we suspect that this result could be contingent on the level of differentiation of the study system. Our very brief analysis of ghost population impacts must be regarded as very preliminary in any general sense. However, the problem is very poorly researched and we hope that our results will stimulate further work on an important issue for modern genetic analysis.

Conclusion: There are two major conclusions from this study. The first is that by blocking compensatory coun-

tercurrent migration anthropogenic barriers have a major impact on *C. gobio* gene flow and the consequent distribution of genetic diversity within the River Rye system. It seems likely that these results will hold not only for bullheads in other rivers systems but also for other headwater riverine species, since all will encounter the ubiquitous force of near unidirectional river flow to some extent. While further investigation is desirable, the potential negative consequences of anthropogenic barriers for population persistence seem clear. We therefore suggest that river managers carefully balance the utility of waterway obstructions against their consequences for conservation of genetic diversity. Second, we found a high degree of concordance among the several estimators of migration and genetic diversity. Even the least congruent estimator—pairwise Nm values calculated from WEIR and COCKERHAM’S (1984) F_{ST} estimator—became far more so if pairwise F_{ST} were used as an index of *relative* gene flow, rather than being translated into Nm . Such agreement between methods suggests that traditional, maximum-likelihood, and nonequilibrium estimators of diversity and migration parameters all performed well for bullheads exhibiting complex natural population structure, deviating in manifold ways from simplified theoretical models. In addition to strengthening our conclusion, we concur with NEIGEL’S (2002) view that it remains important to apply traditional and modern estimators, both to provide a comparative link with past work and to facilitate tests of computationally intensive modern methods, deviations from the assumptions of which are far less well understood than those for F_{ST} .

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LITERATURE CITED

- ABDO, Z., K. A. CRANDALL and P. JOYCE, 2004 Evaluating the performance of likelihood methods for detecting population structure and migration. *Mol. Ecol.* **13**: 837–851.
- BEERLI, P., 2004 Effect of unsampled populations on the estimation of population sizes and migration rates between sampled populations. *Mol. Ecol.* **13**: 827–836.
- BEERLI, P., and J. FELSENSTEIN, 1999 Maximum-likelihood estimation of migration rates and effective population numbers in two populations using a coalescent approach. *Genetics* **152**: 763–773.
- BEERLI, P., and J. FELSENSTEIN, 2001 Maximum likelihood estimation of a migration matrix and effective population sizes in n subpopulations by using a coalescent approach. *Proc. Natl. Acad. Sci. USA* **98**: 4563–4568.
- BONNET, E., and Y. VAN DE PEER, 2002 *z*: a software tool for simple and partial Mantel tests. *J. Stat. Software* **7**: 1–12.
- BOSSART, J. L., and D. P. PROWELL, 1998 Genetic estimates of population structure and gene flow: limitations, lessons and new directions. *Trends Ecol. Evol.* **13**: 202–206.

- BROOK, B. W., D. W. TONKYN, J. J. Q'GRADY and R. FRANKHAM, 2002 Contribution of inbreeding to extinction risk in threatened species. *Conserv. Ecol.* **6**(1): 16. (<http://www.ecologyandsociety.org/vol6/iss1/art16>)
- BRUFORD, M. W., O. HANOTTE, J. F. Y. BROOKFIELD and T. BURKE, 1992 Single locus and multilocus DNA fingerprinting, pp. 227–229 in *Molecular Genetic Analysis of Populations: A Practical Approach*, edited by A. R. HOELZEL. IRL Press, Oxford.
- CIOFI, C., M. A. BEAUMONT, I. R. SWINGLAND and M. W. BRUFORD, 1999 Genetic divergence and units for conservation in the Komodo dragon *Varanus komodoensis*. *Proc. R. Soc. Lond. Ser. B Biol. Sci.* **266**: 2269–2274.
- CORANDER, J., P. WALDMANN, P. MARTTINEN and M. J. SILLANPAA, 2004 BAPS 2: enhanced possibilities for the analysis of genetic population structure. *Bioinformatics* **20**: 2363–2369.
- CORNUET, J. M., and G. LUIKART, 1996 Description and power analysis of two tests for detecting recent population bottlenecks from allele frequency data. *Genetics* **144**: 2001–2014.
- DOWNHOWER, J. F., P. LEJEUNE, P. GAUDIN and L. BROWN, 1990 Movements of the Chabot (*Cottus gobio*) in a small stream. *Pol. Arch. Hydrobiol.* **37**: 119–126.
- EFREMOV, V. V., 2004 The rate of approach to the equilibrium value of F-ST in island and one-dimensional stepping-stone models of migration. *Russ. J. Genet.* **40**: 1041–1045.
- ENGBRECHT, C. C., C. R. LARGIADER, B. HÄNFLING and D. TAUTZ, 1999 Isolation and characterization of polymorphic microsatellite loci in the European bullhead *Cottus gobio* L. (Osteichthyes) and their applicability to related taxa. *Mol. Ecol.* **8**: 1966–1969.
- FAGAN, W. F., 2002 Connectivity, fragmentation, and extinction risk in dendritic metapopulations. *Ecology* **83**: 3243–3249.
- FAGAN, W. F., C. AUMANN, C. M. KENNEDY and P. J. UNMACK, 2005 Rarity, fragmentation, and the scale dependence of extinction risk in desert fishes. *Ecology* **86**: 34–41.
- FRASER, D. J., C. LIPPE and L. BERNATCHEZ, 2004 Consequences of unequal population size, asymmetric gene flow and sex-biased dispersal on population structure in brook charr (*Salvelinus fontinalis*). *Mol. Ecol.* **13**: 67–80.
- GARZA, J. C., and E. G. WILLIAMSON, 2001 Detection of reduction in population size using data from microsatellite loci. *Mol. Ecol.* **10**: 305–318.
- GEYER, C., 1994 *Estimating Normalizing Constants and Reweighting Mixtures in Markov Chain Monte Carlo*. Technical report no. 568 R(4). University of Minnesota, Minneapolis.
- GEYER, C., and E. A. THOMPSON, 1994 Annealing Markov chain Monte Carlo with Applications to Ancestral Inference. Technical report no. 589 R(1). University of Minnesota, Minneapolis.
- GOUDET, J., 2001 *FSTAT: A Program to Estimate and Test Gene Diversities and Fixation Indices*, Version 2.9.3. <http://www.unil.ch/izea/software/fstat.html>.
- HÄNFLING, B., and R. BRANDL, 1998a Genetic differentiation of the bullhead *Cottus gobio* L. across watersheds in central Europe: evidence for two taxa. *Heredity* **80**: 110–117.
- HÄNFLING, B., and R. BRANDL, 1998b Genetic variability, population size and isolation of distinct populations in the freshwater fish *Cottus gobio* L. *Mol. Ecol.* **7**: 1625–1632.
- HÄNFLING, B., B. HELLEMANS, F. A. M. VOLKAERT and G. R. CARVALHO, 2002 Late glacial history of the cold-adapted freshwater fish *Cottus gobio*, revealed by microsatellites. *Mol. Ecol.* **11**: 1717–1729.
- HERSHEY, A. E., J. PASTOR, B. J. PETERSON and G. W. KLING, 1993 Stable isotopes resolve the drift paradox for baetis mayflies in an arctic river. *Ecology* **74**: 2315–2325.
- HUMPHRIES, S., and G. D. RUXTON, 2002 Is there really a drift paradox? *J. Anim. Ecol.* **71**: 151–154.
- HUTCHISON, D. W., and A. R. TEMPLETON, 1999 Correlation of pairwise genetic and geographic distance measures: inferring the relative influences of gene flow and drift on the distribution of genetic variability. *Evolution* **53**: 1898–1914.
- KIMURA, M., and G. H. WEISS, 1964 The stepping stone model of population structure and the decrease of genetic correlation with distance. *Genetics* **49**: 561–576.
- KINNISON, M. T., P. BENTZEN, M. J. UNWIN and T. P. QUINN, 2002 Reconstructing recent divergence: evaluating nonequilibrium population structure in New Zealand chinook salmon. *Mol. Ecol.* **11**: 739–754.
- KNAEPKENS, G., L. BRUYNDONCX, L. BERVOETS and M. EENS, 2002a The presence of artificial stones predicts the occurrence of the European bullhead (*Cottus gobio*) in a regulated lowland river in Flanders (Belgium). *Ecol. Freshwater Fish* **11**: 203–206.
- KNAEPKENS, G., D. KNAPEN, L. BERVOETS, B. HÄNFLING, E. VERHEYEN *et al.*, 2002b Genetic diversity and condition factor: a significant relationship in Flemish but not in German populations of the European bullhead (*Cottus gobio* L.). *Heredity* **89**: 280–287.
- KNAEPKENS, G., L. BRUYNDONCX and M. EENS, 2004 Assessment of residency and movement of the endangered bullhead (*Cottus gobio*) in two Flemish rivers. *Ecol. Freshwater Fish* **13**: 317–322.
- LUIKART, G., and J. M. CORNUET, 1998 Empirical evaluation of a test for identifying recently bottlenecked populations from allele frequency data. *Conserv. Biol.* **12**: 228–237.
- MELDGAARD, T., E. E. NIELSEN and V. LOESCHCKE, 2003 Fragmentation by weirs in a riverine system: a study of genetic variation in time and space among populations of European grayling (*Thymallus thymallus*) in a Danish river system. *Conserv. Genet.* **4**: 735–747.
- MORAN, M. D., 2003 Arguments for rejecting the sequential Bonferroni in ecological studies. *Oikos* **100**: 403–405.
- MORITA, K., and S. YAMAMOTO, 2002 Effects of habitat fragmentation by damming on the persistence of stream-dwelling charr populations. *Conserv. Biol.* **16**: 1318–1323.
- MORITA, K., and A. YOKOTA, 2002 Population viability of stream-resident salmonids after habitat fragmentation: a case study with white-spotted charr (*Salvelinus leucomaenis*) by an individual based model. *Ecol. Modell.* **155**: 85–94.
- MÜLLER, K., 1954 Investigations on the organic drift in North Swedish streams. *Inst. Freshwater Res. Rep.* **34**: 133–148.
- NEI, M., 1978 Estimation of average heterozygosity and genetic distance from a small number of individuals. *Genetics* **89**: 583–590.
- NEIGEL, J. E., 2002 Is F-ST obsolete? *Conserv. Genet.* **3**: 167–173.
- NOLTE, A. W., J. FREYHOF, K. C. STEMSHORN and D. TAUTZ, 2005 An invasive lineage of sculpins, *Cottus* sp. (Pisces, Teleostei) in the Rhine with new habitat adaptations has originated from hybridization between old phylogeographic groups. *Proc. R. Soc. Lond. B Biol. Sci.* **272**: 2379–2387.
- PACHEPSKY, E., F. LUTSCHER, R. M. NISBET and M. A. LEWIS, 2005 Persistence, spread and the drift paradox. *Theor. Popul. Biol.* **67**: 61–73.
- PAETKAU, D., R. SLADE, M. BURDEN and A. ESTOUP, 2004 Genetic assignment methods for the direct, real-time estimation of migration rate: a simulation-based exploration of accuracy and power. *Mol. Ecol.* **13**: 55–65.
- PIRY, S., G. LUIKART and J. M. CORNUET, 1999 BOTTLENECK: a computer program for detecting recent reductions in the effective population size using allele frequency data. *J. Hered.* **90**: 248–249.
- PIRY, S., A. ALAPETITE, J. M. CORNUET, D. PAETKAU, L. BAUDOUIN *et al.*, 2004 GENECLASS2: a software for genetic assignment and first-generation migrant detection. *J. Hered.* **95**: 536–539.
- RANNALA, B., and J. L. MOUNTAIN, 1997 Detecting immigration by using multilocus genotypes. *Proc. Natl. Acad. Sci. USA* **94**: 9197–9201.
- RAYMOND, M., and F. ROUSSET, 1995 GENEPOP (version 1.2): a population genetic software for exact tests and ecumenicism. *J. Hered.* **86**: 248–249.
- ROUSSET, F., 1996 Equilibrium values of measures of population subdivision for stepwise mutation processes. *Genetics* **142**: 1357–1362.
- SACCHERI, I., M. KUUSAAARI, M. KANKARE, P. VIKMAN, W. FORTELIUS *et al.*, 1998 Inbreeding and extinction in a butterfly metapopulation. *Nature* **392**: 491–494.
- SLATKIN, M., 1993 Isolation by distance in equilibrium and non-equilibrium populations. *Evolution* **39**: 53–65.
- SLATKIN, M., 2005 Seeing ghosts: the effect of unsampled populations on migration rates estimated for sampled populations. *Mol. Ecol.* **14**: 67–73.
- SLATKIN, M., and N. H. BARTON, 1989 A comparison of 3 indirect methods for estimating average levels of gene flow. *Evolution* **43**: 1349–1368.
- SPEIRS, D. C., and W. S. C. GURNEY, 2001 Population persistence in rivers and estuaries. *Ecology* **82**: 1219–1237.
- VAN HOUTD, J. K. J., J. PINCEEL, M. C. FLAMAND, M. BRIQUET, E. DUPONT *et al.*, 2005 Migration barriers protect indigenous

- brown trout (*Salmo trutta*) populations from introgression with stocked hatchery fish. *Conserv. Genet.* **6**: 175–191.
- VAN OOSTERHOUT, C., W. F. HUTCHINSON, D. P. M. WILLS and P. SHIPLEY, 2004 MICRO-CHECKER: software for identifying and correcting genotyping errors in microsatellite data. *Mol. Ecol. Notes* **4**: 535–538.
- VITALIS, R., and D. COUVET, 2001a Estimation of effective population size and migration rate from one- and two-locus identity measures. *Genetics* **157**: 911–925.
- VITALIS, R., and D. COUVET, 2001b ESTIM 1.0: a computer program to infer population parameters from one- and two-locus gene identity probabilities. *Mol. Ecol. Notes* **1**: 354–356.
- WAKELEY, J., 2001 The coalescent in an island model of population subdivision with variation among demes. *Theor. Popul. Biol.* **59**: 133–144.
- WEIR, B. S., and C. C. COCKERHAM, 1984 Estimating F-statistics for the analysis of population structure. *Evolution* **38**: 1358–1370.
- WHITLOCK, M. C., and D. E. MCCAULEY, 1999 Indirect measures of gene flow and migration: F-ST not equal $1/(4Nm+1)$. *Heredity* **82**: 117–125.
- WILKINSON-HERBOTS, H. M., and R. ETTRIDGE, 2004 The effect of unequal migration rates on F-ST. *Theor. Popul. Biol.* **66**: 185–197.
- WILLIAMSON-NATESAN, E. G., 2005 Comparison of methods for detecting bottlenecks from microsatellite loci. *Conserv. Genet.* **6**: 551–562.
- WOFFORD, J. E. B., R. E. GRESSWELL and M. A. BANKS, 2005 Influence of barriers to movement on within-watershed genetic variation of coastal cutthroat trout. *Ecol. Appl.* **15**: 628–637.
- WRIGHT, S., 1951 The genetical structure of populations. *Ann. Eugen.* **15**: 323–354.
- YAMAMOTO, S., K. MORITA, I. KOIZUMI and K. MAEKAWA, 2004 Genetic differentiation of white-spotted charr (*Salvelinus leucomaenis*) populations after habitat fragmentation: spatial-temporal changes in gene frequencies. *Conserv. Genet.* **5**: 529–538.
- ZAYKIN, D., L. ZHIVOTOVSKY and B. S. WEIR, 1995 Exact tests for association between alleles at arbitrary numbers of loci. *Genetica* **96**: 169–178.

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